

ORIGINAL ARTICLE

Extraction and Separation of Trace Amounts Lead (II) in Water Samples Using Nano Polyacrylonitril Modified Ethylendiamins

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KEYWORDS

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ABSTRACT: A simple, sensitive and selective method for extraction of trace amounts of Pb (II) in water samples based on the adsorption of its modification nano polyacrylonitrile fiber (PANF) was examined in this study. PANF was prepared by adding acrylic fibers to ethylenediamine (EDA) with different concentration solutions. A new technique using a solid phase extraction (SPE) cartridge with modified PANF as sorbent was developed for the preconcentration of trace amounts of lead and was determined by flame atomic absorption spectrometry (FAAS). Ethylenediamine plays a key role as chelating reagent on ultrahigh specific surface of PANF. Some of the important parameters on the preconcentration and complex formation were selected and optimized. Under the optimized conditions the limit of detection (LOD) and limit of quantification (LOQ) were 0.167, 0.562 and the proposed method has a good reproducibility 0.81% (RSD %). The enrichment factor was 200 and the percentage of recovery was in the range of 95-100%. The method was successfully applied to the recovery of Pb²⁺ in different type of water samples. PANF and its derivate such as PANF-EDA in this study are full of potential to use as an excellent adsorbent in the extraction method like SPE and solid phase micro extraction (SPME).

INTRODUCTION

Though Pb (II) ions are really important for several biological processes, its levels in some states are representation of a problem or illness in many living

systems or environments. These elements accumulate in living organisms and are of high toxic potential [1]. Several analytical techniques such as flame atomic

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absorption spectrometry (FAAS) are available for the determination of trace metals with enough sensitivity for the most applications [2]. FAAS with its relative low cost and decent analytical performance is the important instrument in the research laboratories for determination of a range of heavy metals. Accurate determination of trace heavy metals by FAAS is usually one of the important problems for the analytical chemist because of their low concentrations. Likewise other important problem in FAAS determinations of heavy metals are the effects of the matrix of the analyzed samples. In order to overcome, these problem analytical chemists mostly use separation and preconcentration methods such as coprecipitation, liquid–liquid extraction (LLE), cloud point extraction (CPE), electrodeposition, solid phase extraction (SPE), solid phase micro extraction (SPME), etc. [3-5].

In recent years, SPE method has been well used for determination of lead ions in numerous environmental samples [6, 7] because of its simplicity, rapidity, minimal cost, low consumption of reagents and its ability to combine with different detection techniques either in on-line or off-line mode [8]. The main part of the SPE is the sorbent material that determines the selectivity and sensitivity of the technique. However, the commonly used SPE sorbents, such as C₁₈ silica and graphitic carbon, are often the only suitable option for a limited number of analytes. Reusability of the SPE column is also a problem. Thus, developing new SPE sorbent material is important [9]. Carbon derivatess are well known for their high adsorption capacity. They have been established to possess excessive potential as adsorbents for removing many types of environmental pollutants such as heavy metals[10, 11]: carbon nanotubes (CNTs) [12, 13], Fullerenes can be used as chromatographic stationary phases to offer high selectivity for specific compounds [14, 15] or as sorbent materials for on-line clear up and preconcentration [16, 17].

Recently, CNTs have been excellent classes of sorbent materials for SPE [18]. Since the first application of CNTs in SPE by Cai et al. that he used multi-walled carbon nanotubes (MWCNTs) [18], in recent years many reports have been released focusing on progress of use CNTs-based SPE methods for a great variety of analytes, including phenolic compounds [19], insecticides [20], pharmaceuticals [21, 22], inorganic ions [23], organometallic compounds [24]. The selectivity of adsorbent (CNTs) can be controlled by covalently or non-covalently modification of the CNTs functional groups. In addition, inherent properties of CNTs such as well chemical and thermal stability make them appropriate to be used as adsorbents for SPE column. Other carbon allotropes, for instance, graphite and diamond, have also been established as adsorbents in SPE or SPME [25].

Nevertheless, the sorbents have the advantages of no swelling, fast kinetics, and good mechanical stability as well [22]. Recent research on the use of surfactant-coated mineral oxides for SPE have demonstrated these new sorbent materials to be a promising tool for the preconcentration of organic compounds in a wide polarity range [23]. The most prominent among the supports used are silica gel, [24, 25-28] SDS coated alumina, [29] and modified chromosorb and C₁₈ bonded silica [30].

In this study, we report the synthesis of this new sorbent and its application as a selective sorbent for separation, preconcentration and determination of Pb²⁺ ions by FAAS determination.

MATERIALS AND METHODS

Apparatus

The concentration of the metal ion solutions was determined by using the Varian model spectra AA-240 (Mulgrara, Victoria, Australia). The pH-measurements of the metal ion and buffer solutions were carried out by an Orion 420. Infrared spectra of PANF were carried out

from KBr by a Perkin-Elmer 1430 ratio recording spectrophotometer.

Materials and reagents

All the necessary materials and reagents were of analytical grade and were purchased Merck, Aldrich and Sigma Company. All the dilutions were prepared by ultrapure deionized water.

Preparation of Modified nano polyacrylonitrile fiber

Modified PANF was prepared by adding 3 g of acrylic fibers to 300 ml of Ethylenediamin (EDA) with different concentration solutions. Electro Spinning is done in according to reference [33]. The adsorption ions onto PANF-EDA for Pb (II) ions were investigated using the batch method [28].

Column preparation and recommended procedures

The adsorption of lead by the adsorbents was studied by the SPE technique. The SPE mode adsorption was selected because of its simplicity, fast and economic for the preconcentration and determination of trace amounts of lead in different samples. PANF-EDA (0.1 g) was placed in a 2 mL SPE column using an upper frit and a lower frit to avoid adsorbent loss. Prior to extraction, the column was conditioned with 10mL of MeOH and 10mL of deionized water, respectively. The sample solution (50mL) was passed through the column at a flow rate of 2mL/min. Then, the column was washed with 5mL of 10% (v/v) MeOH aqueous solution to remove the co-adsorbed matrix materials from the column. The analytes retained on the column were eluted with 5mL of 4M HNO₃ aqueous solution that flow rate of eluent was 2ml/min.

Finally, the analyte ions in the eluent were determined by FAAS at 228.8 nm for Pb. The effects of several parameters, such effect of pH, eluent type and its volume, effect of flow rates of sample and eluent solution and amount of adsorbent, effect of interfering

ions and breakthrough volume were also studied. The results of these studies were used to obtain the optimum conditions for adsorption capacity measurements. Using the procedure described above, the percent of recovery was calculated from the following equation) Eq. (1):

$$R\% = \frac{(C_s \cdot V_s) \times \left(\frac{V_w}{V_s}\right)}{V_w \cdot C_w} \quad (1)$$

Where C_s is analyte concentration in eluent (found by FAAS) (mg/L), and C_w is analyte concentration in sample solution (known) (mg/L), V_s is volume of eluent (mL) and V_w is volume of sample solution (mL).

Sampling

Tap water(Tehran, taken after 10 min operation of the tap), rain water (Tehran, 20 January, 2014), and Sea water (taken from Caspian sea, near the Mahmoud-Abad shore) samples were analysed (Table 2). Tap water samples used for development of the method were collected in glasses containers. Before the analysis, the organic content of the water samples was oxidized in the presence of 7% H₂O₂ and then concentrated nitric acid was added. These water samples were then filtered through a 0.45µm Millipore cellulose membrane to remove suspended particulate matter and stored in a refrigerator at 4 °C in the dark before analysis.

RESULTS AND DISCUSSION

Some preliminary experiments were carried out in order to investigate the extraction of Pb by the PANF-EDA from solution. The results showed that PANF-EDA could extract it quantitatively. Figure 1 shows SEM image of TEM images of the nano polyacrylonitrile fiber.

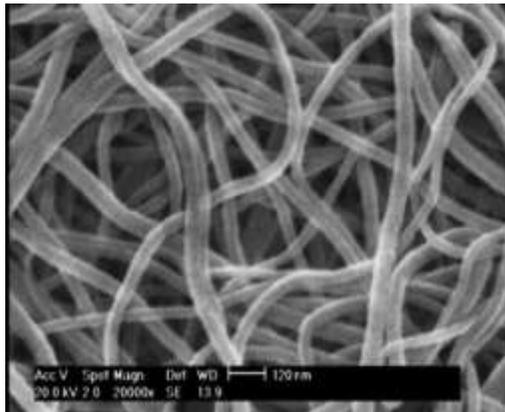


Figure 1. SEM image of TEM images of the nano polyacrylonitrile fiber

SEM Investigations

SEM was to examine the external surface of the fiber before and after the modification. As can be seen from Figure 2, acrylic fiber similarly surface (Figure 2 (a)), and with modified fiber (PANF-EDA), obvious change comparing to that of the RAF fiber was observed (Figure 2 (b)). It is clear that the changes have occurred in the morphology of the fiber but photographs demonstrated that the surface of PANF-EDA was approximately as smooth, swollen and homogeneous as that of the raw fiber. This can be related to new functional groups that were bigger than (CN) groups.

SEM was examining the morphology of the nano fiber before and after the modified. As can be seen from Figure 3, original acrylic nano fiber comparatively morphology (Figure 3(a)), and with modification nano fiber (PAN-EDA), obvious change compared to that of the unmodified fiber was observed (Figure 3 (b)). The PAN-EDA was roundelay as that of unmodified fiber acrylic nano fiber. This can be related to the modified treatment and incorporation of new functional groups into the fiber structure

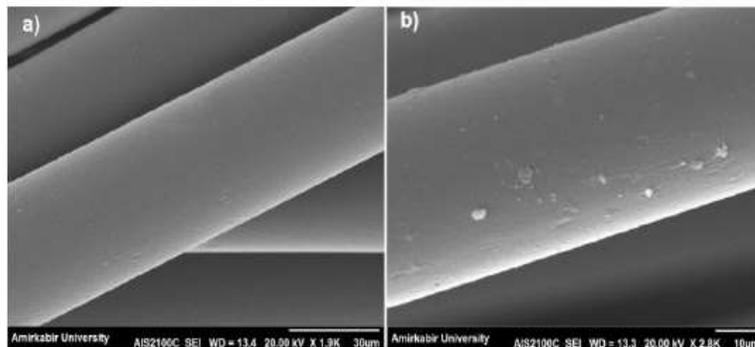


Figure 2. SEM image of (a) the raw fiber and (b) modified PAN fiber

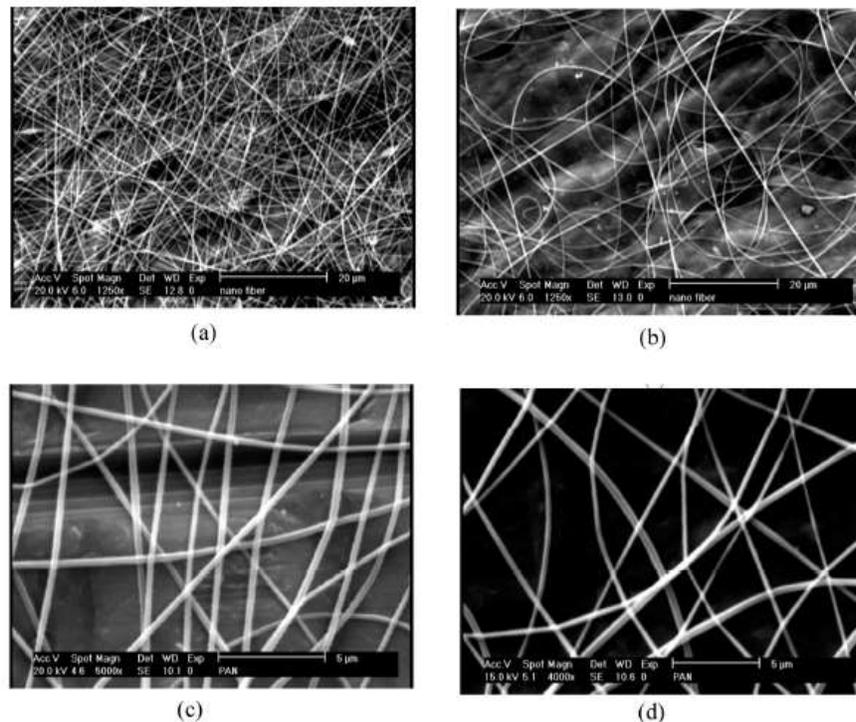


Figure 3. SEM image of (a) (c) the raw nano fiber and (b)(d) modified PAN nano fiber

Optimization of SPE procedures

Effect of pH

The metal chelate stability constant and its chemical stability considerably influence the SPE recovery. The pH plays a very important role on metalchelate formation and following extraction. Therefore, pH was the first optimized parameter. pH of the analyte solutions was adjusted to desired values with diluted hydrochloric acid (0.1 mol L^{-1}) and/or ammonia solution (0.1 mol L^{-1}) The variation in recovery of Pb(II) with pH is shown in Figure 4. According to the results shown in Figure 4 up to pH 3-3.5, complete

recoveries are obtained. However, at higher pH values, percentage recovery decreases. This is due to fact that in an acidic solution the protonation of PANF-EDA occurs and there is a weak tendency for retention between Pb (II) and PANF-EDA, whereas at higher values ($\text{pH} > 4$), Pb (II) reacts with hydroxide ions to produce $\text{Pb}(\text{OH})_2$. Accordingly, pH 3.0 was selected for subsequent work and real sample analysis.

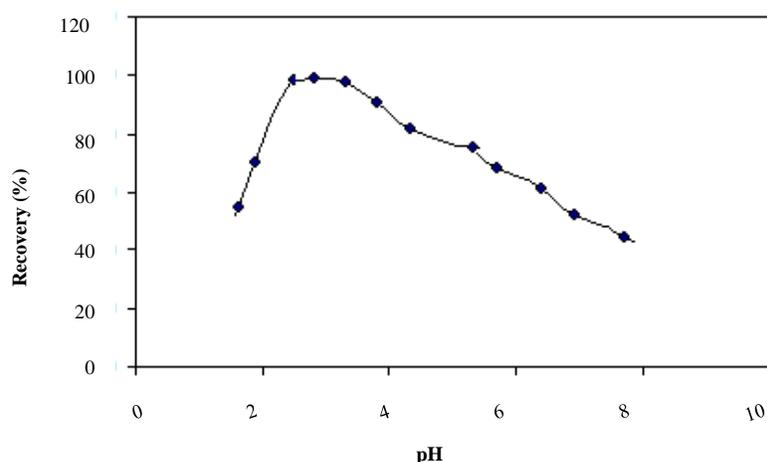


Figure 4. Effect of pH on the extraction of Pb. Experimental conditions: source, 10 ml of $0.1\mu\text{gml}^{-1}$ Pb^{2+} solution; amount of sorbent, 50.0 mg

Effect of flow rates of sample and eluent solution

The effects of flow rates of sample solution and eluent solution on the recovery of Pb (II) was examined between under the optimum conditions in the range of $1\text{--}10.0\text{ mL min}^{-1}$ by controlling the flow rate with peristaltic pump. The recovery of the ions were independent of flow rate in the range of $0.5\text{--}2.0\text{ mL min}^{-1}$ for eluent solution and range of $1\text{--}5.0\text{ mL min}^{-1}$ for sample solution.

Effect of sample solution volume

Another parameter studied to find the best experimental conditions is the volume of sample solution and/or analyte concentration. For this purpose, $50.0\text{--}1100.0\text{ mL}$ of sample solutions containing 5 ppm Pb (II) was processed according to the suggested procedure. The recovery of Pb(II) was quantitative ($>96\%$) obtained up to a sample volume of 1000.0 mL and the adsorbed

Pb(II) can be eluted with 5 mL eluent. Therefore, an enrichment factor of 200 was achieved by this technique. Finally In our suggested procedure, a sample volume of 50.0 mL was chosen for preconcentration method.

Effect of amount of sorbent

To achieve a high extraction recovery, different amounts of PANF-EDA ranging from 50 to 300 mg were applied to extract the target compounds from the sample solutions. The results are shown in Figure 5 from which it can be seen that the extraction recovery achieved by 100 mg, but almost the same as obtained with 300 mg or more than of the adsorbent. Based on the above results, 100 mg of PANF-EDA was selected for the following experiments.

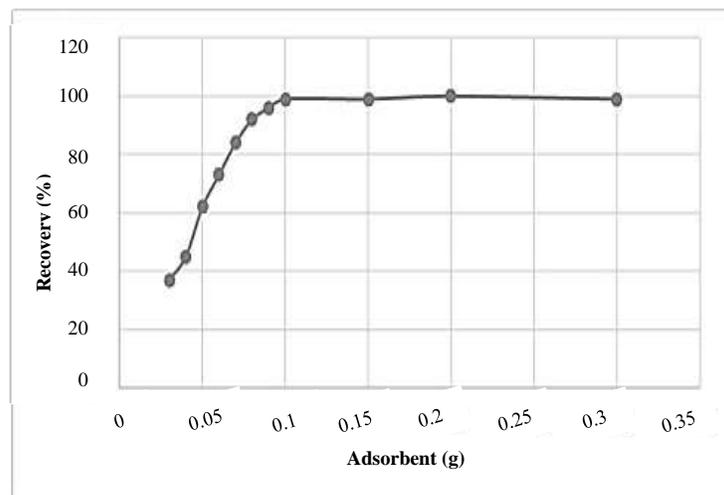


Figure 5. Removal percentages of Pb (II) at different amounts of adsorbent, of Pb (II), pH= 3.0.

Eluent type and its volume

Other important factors, which affect the percent of recovery, are the type, volume, and concentration of the eluent solution used for the removal of metal ions from the sorbent. For this purpose, various types of eluents were examined according to the suggested procedure. Four mol L⁻¹ HNO₃ was found to be adequate for quantitative elution (≥95%). The results show the best quantitative recovery is 5.0 mL of 4.0 mol L⁻¹ HNO₃. As a result, 5.0 mL of 4.0 mol L⁻¹ HNO₃ was selected in the subsequent preconcentration method.

Reusability of Column

The stability and potential regeneration of the column were studied. After every time extraction, the column was washed with 10 mL of MeOH and 10 mL of deionized water. Thus, the column was available for a

next extraction immediately at least 50 adsorption elution cycles without significant decrease in the recovery of Pb (II) ions.

Effect of foreign ions

The influence of common foreign ions on the adsorption of Pb (II) on PANF-EDA was studied. In these work, 50.0 ml solutions containing 5 ppm of Pb and various amounts of interfering ions were treated according to the suggested procedure. The tolerance level was defined as the maximum concentration of the foreign ion causing a change in the analytical signal no higher than 5%, when compared with the signal of 5-ppm lead alone. The results, listed in Table 1, demonstrate that the presence of major cations and anions in natural water has no important influence on the adsorption of Pb (II) ions under the designated conditions.

Table 1. Separation of Pb from binary mixtures ^a

Diverse ion	Amounts taken(mg)	% Found	%Recovery of Pb ²⁺ ion
Na ⁺	92.4	1.15(2.4) ^b	98.5(2.7)
K ⁺	92.5	1.36(2.3)	98.0(2.2)
Mg ²⁺	24.5	0.70(2.6)	98.5(1.7)
Ca ²⁺	26.3	2.65(3.0)	98.5(1.8)
Sr ²⁺	2.45	2.85(2.1)	98.4(2.0)
Ba ²⁺	3.66	3.16(2.1)	98.7(2.3)
Mn ²⁺	2.66	1.75(2.2)	96.3(2.3)
Co ²⁺	2.17	6.44(2.3)	93.0(1.9)
Cu ²⁺	1.64	2.43(2.4)	93.7(2.4)
Zn ²⁺	2.76	4.97(2.1)	97.6(2.4)
Cd ²⁺	2.77	2.96(2.4)	97.2(2.7)
Hg ²⁺	1.67	2.71(2.1)	97.7(2.7)
Ag ⁺	2.6i	3.47(2.9)	97.6(2.3)
UO ²⁺	2.76	2.74(2.1)	98.3(2.7)

^aInitial samples contained 10µg Pb²⁺ and different amounts of various ions in 100 mL water.

^bValues in parentheses are RSDs based on five individual replicate analysis.

Analytical figures of merits

Under optimized conditions, a calibration curve for Pb (II) was found by preconcentrating a series of Pb (II)

standards according to the procedure mentioned. The curve was linear from 1.0 mg/l to 7.0 mg/l for Pb.

Table 2. Recovery of Pb (II) added to 100mL of different water samples (containing 0.1M buffer acetic acid / acetate at pH= 3.0).

Sample	Pb ²⁺ spiked (ng ml ⁻¹)	Pb ²⁺ detected (ng ml ⁻¹)	Recovery (%)
Sample 1 ^a	5.0	4.8 (2.0) ^b	97.5
	10.0	9.7 (2.5)	98.4
Sample 2 ^c	5.0	4.4 (3.0)	98.3
	10.0	9.2 (3.0)	97.2
Tap water ^d	0.0	2.5 (2.4)	98.5
	5.0	7.5 (2.9)	97.5
	10.0	12.3 (2.4)	98.3
Rain water ^f	0.0	N.D.	—
	5.0	4.9 (2.3)	96.4
	10.0	10.1 (2.6)	97.6
Sea water ^g	0.0	14.0(2.0)	97.3
	5.0	18.9 (2.3)	98.9
	10.0	24.0 (3.0)	98.9

^a Hg²⁺, Co²⁺, Fe³⁺, Ni²⁺, Cr³⁺, 5000 ng ml⁻¹ of each cation; K⁺ and Li⁺, 10,000 ng ml⁻¹ of each, ^b R.S.D of three replicate experiments.

^c Hg²⁺, Co²⁺, Fe³⁺, Ni²⁺, Cr³⁺, 2500 ng ml⁻¹ of each cation; K⁺ and Li⁺, 5000 ng ml⁻¹ of each., ^d From drinking water system of Tehran.

^e Not detected. ^f Tehran, 20 January, 2014, Iran ^g Caspian sea water.

Table 3. Comparison of some methods for preconcentration and determination of thal-lium with proposed method

Method	DL (ng ml ⁻¹)	LDR (ng ml ⁻¹)	Reference
Solid-liquid extraction	1	3.75–17.5	[30]
Potentiometric	0.08	0.1–100	[31]
stripping	1	5–250	[32]
Liquid-liquid extraction Microcrystalline naphthalene	4	5–20	[33]
	20	40–180.0	[34]
Proposed method	0.167	5.0–240.0	–

Determination of Lead in real water samples

Three type of water samples (information described in section 2.6 sampling) were used for the determination of lead. The analytical results are given in Table 2. The percent of recoveries for the addition of different concentrations of Pb (II) to water samples were 97 and 98.9%. These satisfactory percent of recoveries indicate no significant effects from the matrix composition of the real water samples.

Comparison with other solid phase adsorbents

The proposed methodology was compared to a variety of solid adsorbents reported recently in the literature. The distinct features are summarized in Table 3. As can be seen from the table, it is evident that the preconcentration factor obtained with alumina is comparable to or even better than most of the other chelating matrices. The other significant feature of the proposed method is the use of environmentally.

CONCLUSIONS

The proposed SPE possesses the following advantages: the technique is rapid when compared with the previously reported procedures for the separation and determination of lead; the time taken for the separation and determination of lead in a 500 mL sample is at the most 30 min.

The proposed method reveals the great potential of PANF-EDA as an advantageous sorbent material in SPE. Using Pb²⁺ as model analyte, the PANF-EDA packed SPE columns showed reliable and attractive

analytical performance in the analysis of environmental water samples. Higher recoveries were achieved with PANF-EDA than with other adsorbents including C₁₈ silica, graphitic carbon, and CNTs, owing to the large surface area and unique chemical structure of PANF-EDA. Some other advantages of PANF-EDA as an SPE adsorbent have also been demonstrated, such as high sorption capacity, good reusability, and fine reproducibility. Although the obtained results of this research were related to the Pb²⁺ determination, the system could be a considerable potential guide for the preconcentration and determination of other metals.

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